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2	Greenhouse gas balance as affected by no-tillage and winter cover crops in a	
3	subtropical paddy rice ecosystem	
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24 Greenhouse gas balance as affected by no-tillage and winter cover crops in a

25 subtropical paddy rice ecosystem

26

27 Abstract

28 Paddy rice production based on traditional soil management practices emits large 29 amounts of methane (CH_4) to the atmosphere. Here, we assessed the The potential of no-30 tillage (NT) and winter cover crops (WCC) to mitigate net greenhouse gas (GHG) 31 emissions and yield-scaled GHG emissions in a subtropical paddy rice ecosystem was investigated in a Gleysol of Southern Brazil. A long term (22-yrs) experiment was used 32 33 to evaluated evaluate regarding the effect of NT combined with winter fallow or three WCC (ryegrass, white oat and birdsfoot trefoil) on seasonal CH₄-C and nitrous oxide 34 35 (N₂O-N) emissions and on soil organic C-carbon (SOC) stocks in comparison to 36 conventional tillage (CT) under winter fallow. in a Gleysol of Southern Brazil. The 37 changes in SOC were used as a proxy for annual net carbon dioxide (CO₂) exchanges in 38 the soil-atmosphere. taking tThe CT soil was taken as a reference, and the GHG balance 39 (summation of CH₄, N₂O and CO₂ emissions multiplied by their global warming potential 40 of 34, 298 and 1, respectively) and yield-scaled GHG emissions (GHG emissions per unit 41 of grain yield) were calculated. Across-Among_winter managementsmanagement 42 practises, NT decreased reduced 25% of total GHG emissions by 25% in comparison tocompared with the CT system. This effect was mainly related to decrease of seasonal 43 44 CH_4 -C emissions (31-113 kg ha⁻¹) and C sequestration in soil (0.45-0.65 Mg ha⁻¹ yr⁻¹) in 45 comparison to CT system, since soil N₂O-N emission was not affected by management practices. The sensible? decrease of CH4-C emissions in NT, even under winter fallow 46 47 with low biomass input, suggest that, in addition to the maintenance of crop residues on 48 the soil surface, the lower emissions in NT soil were also related to the lower exposure of

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soil labile-C to methanogenic microorganisms. WCC did not determine-provide an
additional decrease on in-_net GHG emissions in-from NT soil when compared to-with
fallow due to due to_ the increment-change in soil C sequestration resulting from higher
plant biomass input was-being partially off-set by increased soil CH₄-C emissions. Based
on our findings, NT mitigates net GHG emissions in subtropical paddy rice ecosystems,
but no additional effect is observed if combined with WCC.

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56 Keywords: nitrous oxide, methane, soil tillage, carbon sequestration, wetlands

57 Introduction

The development of agricultural production systems that allow, at the same time,simultaneously produce the production of healthy food and minimise mitigation of greenhouse gases-gas_(GHG) emissions is one of the major challenges facing global warming and climate change (Princiotta 2009; Peters *et al.* 2013). Agriculture and land use change contribute to 25% of total GHG emission at the global level (Smith *et al.* 2014), while in Brazil this contribution increase amounts to 51% due to deforestation and a large agricultural crop area (Brasil 2016).

65 The agricultureAgricultural practises developed in the Brazilian lowlands, 66 especially particularly the irrigated rice, has a particular characteristically of high soil 67 methane (CH₄) emissions, which contribute contributing to approximately 18% of the total GHG emissions from agriculture in southern Brazil (Observatório do Clima 2019). 68 69 Since soil CH4 fluxes are driven by the balance between methanogenesis and 70 methanotrophy (oxidation of CH_4) (Le Mer and Roger 2001), the anoxic soil condition of the lowlands favours the methanogenesis process and, consequently, CH₄ emission. The 71 72 strong relatively high rate of CH₄ emissions in the lowlands at of this subtropical 73 ecosystem is generally considered to be caused by occurs by the wide adoption of **Commented [A4]:** Can't start a sentence with an abbreviation

74 conventional tillage (CT) systems. These practises with involve high considerable soil disturbance by through ploughing, disking discing and levelling, and the which determine 75 incorporation incorporation of weeds and crop residues into soil which stimulating 76 stimulates CH₄ production (Wang et al. 1998; Zschornack et al. 2011; Bayer et al. 2014). 77 78 In addition to CH₄, soil losses of organic C under conventional tillage CT and low 79 C input resulting offrom winter fallow (Rosa et al. 2011; Ghimire et al. 2011), contribute to net net-emissions of carbon dioxide (CO₂) to atmosphere at from this subtropical 80 ecosystem. While lowland soils under conventional tillageCT are great significant 81 sources of CH₄ and CO₂ flux to the atmosphere, nitrogen fertilization and soil N cycling 82 83 also contribute to nitrous oxide (N2O) emissions from in-rice fields (Zhang et al. 2015; 84 Islam et al. 2018),). but Although its N₂O emission rates from lowland rice are usually lower than those observed in uplands due to.....(Bayer et al. 2014; Bayer et al. 2016). 85

86 A feasible strategy to GHG mitigation is the adoption of no-tillage (NT) that, due to maintaining crop residues on the soil surface, favour a decrease of soil CH₄ emissions 87 (Bayer et al. 2014; Hao et al. 2016) and retention of atmospheric CO₂-C in soil organic 88 89 matter in the less oxidative environment of NT soil (Nascimento et al. 2009; Rui and 90 Zhang 2010). The combination of winter cover crops (WCC) can increase the potential 91 of C sequestration in NT soil (Yagioka et al. 2015; Bayer et al. 2016), but an increase in 92 soil CH_4 emissions is also expected due to crop residues represent providing a labile C source to for the methanogenesis process in soil (Bayer et al. 2014; Haque et al. 2013; 93 94 Zschornack et al. 2011). In the case of legume species of WCC, Nof crop residues residue 95 N cycling in soil may increase mineral N forms and, as a consequence, favour N₂O emission (Zschornack et al. 2011). However, in the Brazilian uplands, this the impact of 96 legume cover crops on N₂O emissions is partially or totally off-set by C sequestration in 97 soil (Bayer et al. 2016). For lowland soils in subtropical ecosystems, there is scant 98

Commented [A5]: Seems a bit contrary to say that cultivating the soil stimulates CH4 production as subsequently ii is stated that NT soil is less oxidative. Maybe it stimulates CH4 release through disturbance, rather than actual production.

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99 knowledge on of the influence of conservation soil management systems comprising NT and WCC on GHG and C sequestration is scant, mainly for subtropical ecosystems. 100 101 Most Many GHG studies in agricultural systems have focused on only one of the 102 three GHGgases (CH₄, N₂O or CO₂). and, thus However, in order to assess, they are not 103 conclusive regarding the real total-impact of management practices on net emissions of 104 GHG, all three gas fluxes. Aiming to have a real outline of the impact of soil management 105 practices on GHG emissions, the balance encompassing the three main GHG emissions 106 must be determined (Robertson et al. 2001; Mosier et al. 2005; Bayer et al. 2016). For 107 calculation of GHG balance, the annual or seasonal emissions of each gas must be is 108 multiplied by their respective radiative forcing index (1 for CO₂, 34 for CH₄ and 298 for 109 N₂O), and then summed (Robertson et al. 2000; Mosier et al. 2005; Bayer et al. 2016). 110 Paired to GHG balance per cultivated area, the-yield-scaled GHG emission measure 111 provides the intensity of GHG per unit of grain yield (Mosier et al. 2006). 112 Our starting hypothesis is that NT and WCC adoption decrease net GHG 113 emissions per area and per unit of yield in subtropical paddy rice ecosystems, what is 114 possibly related to increase of soil C sequestration and mitigation of soil CH4-emission. 115 The main objective of this study was to evaluate the potential of NT and WCC on GHG 116 balance and yield-scaled GHG emissions in a-paddy rice fields in southern Brazil. The 117 hypotheses were i) NT and WCC adoption decrease net GHG emissions per area and per 118 unit of yield in subtropical paddy rice ecosystems. ii) A decrease in net GHG emissions 119 in NT and WCC systems compared with CT is predominantly due to an increase in soil

120 <u>C sequestration and mitigation of soil CH₄ emissions.</u>

122 Material and methods

123 Site description and <u>e</u>Experiment design-and conduction

The study was based on a long-term field experiment (22 years) conducted in Rice Experimental Station of Rio Grandense Rice Institute, in Cachoeirinha, RS, Southern Brazil (29°55'30" S, 50°58'21" O, 7 m a.s.l). The climate is humid subtropical and classified as Cfa, according to the (Köppen classification), with a mean annual temperature and precipitation of 19.3 °C and 1434 mm, respectively (Climate data 2018). The loamy soil (140 g kg⁻¹ clay) is classified as Gleysol according to the FAO (1998).

130 Irrigated rice (Oryza sativa L.) was the summer crop in the whole experiment, and five treatments consisted of different combination of tillage system in spring and cover 131 132 crops or fallow in winter, as follow: (i) conventional tillage (CT) and (ii) no-tillage (NT), 133 both combined with winter fallow, and (iii-v) NT combined with three winter cover crops 134 (WCC) - ryegrass (Lolium multiflorum L.), white oat (Avena sativa L.), and birdsfoot trefoil (Lotus corniculatus L.). The treatments were applied in plots of 5 m × 8.8 m, and 135 136 the distribution of four replicates at field followed were distributed as a complete 137 randomized block design.

The WCC were <u>sowed_sown</u> in April through surface rice crop residues (no-till)
and grown without irrigation. The seed rates for ryegrass, white oat and birdsfoot trefoil
were 80, 20 and 8 kg ha⁻¹, respectively. In treatments with fallow, only weeds and ryegrass
developed in the plots.

Conventional tillage and NT systems were applied in spring. The CT consisted in of ploughing and discking at 25 and 7 cm, respectively. These operations buried the harvest residues of rice and weeds into soil. In <u>the NT system</u>, soil was <u>not-un</u>disturbed and rice, weed and WCC residues were maintained on <u>the soil surface</u>. In <u>the NT system</u>,

weeds control and management of WCC were performed by desiccation with glyphosate
(Roundup[®] 3.5 L p.c. ha⁻¹) about 30 days before rice sowing.

The rice (cultivar IRGA 424 RI) was sowed sown at a seed rate of 100 kg ha-1 in 148 the first half of October every year. The mean rates of P2O5 and K2O applied at rice 149 150 sowing were 108 and 68 kg ha⁻¹. N fertilization rate was applied at 150 kg ha⁻¹ splited in 151 two times, with 66% applied in V4 and 34% in R0 stages (CQFS-RS/SC 2004; SOSBAI 2016). Weed control was carried out in the pre-emergence rice period using clomazone 152 153 herbicide (Gamit® 1.2 L ha⁻¹) and glyphosate (Roundup® 2.0 L ha⁻¹). Post rice emergence, cyhalophope cyhalofop herbicide (Clincher® 1.7 L ha-1) and penoxulan 154 155 penoxsulam (Ricer® 0.25 L ha-1) were used. Pest control was carried out with neonicotinoid-pyrethroid insecticide (Engeo Pleno® 0.2 L ha-1) in R0. Irrigation by 156 157 flooding started in V4 developing development stage of the plants just after the first N 158 application, and a water heigh-t of 5-7 cm was maintained until rice maturation, when the 159 water supply has been cutwas cut off.

160

161 Evaluation of biomass addition by cover crops, weeds and rice

Aboveground biomass of WCC and weeds in winter (2017) and rice in summer (2017/18) were evaluated by sampling an area of 0.5 m² per plot at the beginning of flowering (R4 stage). Biomass samples were oven dried (forced air at 50 °C) until constant mass, and contribution of roots was considered as 30% of shoot addition by WCC or weeds (Zanatta *et al.* 2007) and rice (Insalud *et al.* 2006). Five subsamples were collected per plot.

- 168
- 169 Evaluation of rice yield

170 Rice grain yield evaluation were performed in 2017/2018 crop seasons through a Wintersteiger mechanized harvester and were obtained by extrapolating the yield 171 harvested in the useful area of each plot to one hectare and fitting the grain moisture to 172 130 g kg⁻¹. 173

174

175 Evaluation of CH_4 and N_2O fluxes and calculation of seasonal emissions

Air samples were collected by using the static chamber method (Parkin and 176 177 Venterea 2010), in 2017/2018 growing season. Each chamber consisted of an aluminium base (50 cm × 100 cm) and an aluminium top of the same size. The bases were driven 178 179 0.05 m into the soil before permanent flooding and were left fixed in the throughout the rice season. 180

181 Each base had an open bottom and sealable channels on the sides to facilitate free 182 flowing of irrigation water in the rice season. The channels on the sides were sealed during air sampling events. Each base covered three rows of rice plants. During the rice 183 184 season, an additional 0.20 or 0.30 m aluminium extensions were stacked on the bases as 185 the plants grew up taller and the respective chamber volume was applied for calculation of GHG fluxes. Internally within eEach chamber top had internally there was a rubber 186 187 septum sampling port, and a stainless steel thermometer, and aA battery operated three 188 12V fans for mixing the headspace and homogenize air within the chamber (Bayer et al. 189 2014).

190 Each air sampling session started at 09:00 h, when the aluminium chamber top 191 were was deployed on the aluminium base. Internal air samples were taken with polypropylene syringes (20 mL) at the chamber deployment (time zero) and 10, 20 and 192 193 30 minutes later; and transferred into 12-ml vials previously evacuated (Exetainer®, 194 Labco, UK). The N₂O and CH₄ concentrations were measured in a Shimadzu GC 2014

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chromatograph equipped with a flame ionization detector (FID) for CH₄, and an electron
capture detector (ECD) for N₂O.
The N₂O or CH₄ fluxes were calculated taking into account the: (i) linear increase
of the internal gas concentration during the 30-min deployment, (ii) area covered by the
chamber, (iii) volume, (iv) temperature and (v) pressure (assumed as 1 atm) inside the
chamber (Parkin and Venterea 2010). Assuming that the gases flux from 09:00 to 12:00
represents the average daily flux (Bayer *et al.* 2016), the seasonal emission was calculated

202 by integrating N_2O-N and CH_4-C fluxes during the monitoring period.

203

204 Soil sampling, organic C analysis and calculation of stocks

In September 2015, soil samples of the 0-2.5, 2.5-5.0, 5.0-7.5, 7.5-10, 10-15, 15-205 206 20, 20-30 and 30-40 cm layers were collected before-rice sowing the rice (i.e., 18 years 207 after the experiment was started). Trenches were dug manually to allow the assessment 208 of soil bulk density using the volumetric ring method (Blake and Hartge 1986). In the 209 same layers as those sampled for bulk density assessment, soil samples were collected 210 with a spatula and then air dried and ground to $\leq 2 \text{ mm}$ in a Marconi 330 grinder. Approximately 20 g of soil was further ground to ≤ 0.025 mm in an agate mortar and 211 212 analyzed for C by dry combustion in a Shimadzu VCSH analyzer. SOC stocks were 213 calculated for the whole entire 0-40 cm profile, by using the equivalent soil mass approach (Ellert and Bettany 1995) with CT soil mass as reference. The annual rate of 214 215 soil organic CSOC accumulation was calculated as the ratio of difference between soil 216 organic CSOC stocks in the treatments and the reference system (CT fallow) and the 217 duration time of the experiment when soil was sampled (18 years).

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Commented [A13]: How many individual measurements were made? Can you provide SOC as a % , pH, TN %

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219 Net balance of GHG emissions and yield-scaled GHG emission

220	The net balance of GHG emissions for each management system was calculated				
221	according to the equation 1:				
	Net balance of $GHG = N_2O + CH_4 - \Delta SOC$	(1),			
222					
223	where N_2O and CH_4 are the cumulative annual emissions of N_2O and CH_4 properly	у			

converted into CO_2 equivalent (CO_{2eq}) after considering the global warming potentials (298 for N₂O, 34 for CH₄) and conversion factors; and ΔSOC is the annual change in soil organic earbon (SOC) stock to 40 cm depth, which was assumed as a proxy of the net exchange of CO_2 in the soil-atmosphere interface (Mosier *et al.* 2005).

The Yield-Scaled GHG was calculated by the ratio between the balance of net GHG
emissions and the average grain yield of rice, to infer the intensity of GHG emission (in
Mg CO₂ eq ha⁻¹) per unit (Mg) of rice grain yield.

231

232 Statistical Analysis

233 The results were checked for variance normality and homoscedasticity with the 234 Shapiro-Wilk and Oneill & Matthews tests, respectively, and appropriated data transformation were performed when assumptions were violated. Seasonal CH₄-C and 235 236 N2O-N emissions, balance of GHG, rice grain yield and yield-scaled GHG emissions 237 were submitted to analysis of variance. When significant at the 5% level, the differences among treatments were subjected to Tukey's post-hoc test at the 5% significance level. 238 239 The MIXED procedure was used to compare the effects soil managements-in winter on 240 GHG, soil organic C and rice yield. The statistical procedures used soil winter management as a fixed factor, and blocks and experimental errors as random variables. 241

242 All analyses were performed with SAS® v. 9.4 (Statistical Analysis System Institute,

243 Cary, NC, USA).

244

245 Results

246 Rice grain yield and annual input of biomass

247 The rice grain yield was not influenced by tillage systems nor by WCC (Table 1), ranging from 8.2 Mg ha-1 in CT-fallow to 8.7 Mg ha-1 in NT- birdsfoot trefoil (Table 2). A 248 249 mean rice yield of 8.4 Mg ha⁻¹ was attained across soil management systems (Table 2). Annual biomass input ranged between 10.4 and 16.8 Mg ha⁻¹ (Table 2). In the 250 251 winter, biomass input ranged from 0.7 (CT) to 1.6 Mg ha⁻¹ (NT) in the systems under fallow, while in the systems under cover crops the biomass input increased to 5.4 Mg ha⁻¹ 252 on average, ranging from 3.9 to 6.8 Mg ha⁻¹ (Table 2). The highest biomass input was 253 254 observed for the rice crop, whose which accounted for 86-93% of total annual biomass input in the fallow systems in CT and NT, and for 60-72% for WCC systems. 255

256

257 Seasonal CH_4 and N_2O emissions

Seasonal CH₄-C emissions was affected by management systems (Table 1) and ranged from 359 to 472 kg ha⁻¹ (Table 2). Greater CH₄-C emission was observed in CT with winter fallow compared to NT with winter fallow (<u>a difference of</u> 74 kg CH₄-C ha⁻¹ of difference) (Figure 1a). Under NT, seasonal soil CH₄-C emission was similar comparing from WCC and-fallow, with <u>the</u> exception of white oat <u>whichthat</u>-presented an seasonal soil CH₄-C emission slightly higher (441 kg ha⁻¹) in comparison compared withto the others WCC (366 kg ha⁻¹, on average) (Figure 1a).

Seasonal soil N₂O-N emissions ranged from 0.25 to 0.87 kg ha⁻¹, and were not
influenced by tillage and WCC (Tables 1 and 2, Figure 1b).

267

268 Soil organic C stocks and annual accumulation rates

269 Soil organic carbon content was influenced by the management systems mainly in the surface soil layers (0-2.5 and 2.5-5.0 cm), where NT and WCC favored greater 270 271 SOC content than CT and winter fallow (Figure 2a). Considering the 0-40 cm layer, lower SOC stocks were observed in CT (47.0 Mg ha⁻¹) than NT (53.3 Mg ha⁻¹), when both were 272 273 combined with winter fallow (Figure 2b). Under NT, the increase of biomass input by 274 WCC did not promoted cause an increase in SOC stocks in comparison to winter fallow. In comparison to CT, NT combined with fallow and WCC presented annual accumulation 275 276 rates of SOC ranging from 0.45 to 0.65 Mg ha-1 yr-1 (i.e. 1.7-2.1 Mg CO2eq. ha-1 yr-1; 277 Table 3)

278

279 Net balance of GHG and yield-scaled greenhouse gas emission

Considering the three main GHG (CO₂, CH₄ and N₂O) emitted from the rice field, total GHG emissions ranged between 14.8 and 21.7 Mg CO₂eq. ha⁻¹ among soil management systems (Table 3). In comparison to the traditional management system (CT combined with winter fallow), NT adoption decreased GHG emissions by 5.4 Mg CO₂eq. ha⁻¹ when combined with winter fallow, and by 3.7-6.8 Mg CO₂eq. ha⁻¹ when combined with WCC (Table 3).

Yield-scaled GHG emission was higher in CT (2.6 Mg CO₂ Mg⁻¹ grain rice) than
in NT (1.9 Mg CO₂ Mg⁻¹ grain rice), both combined with winter fallow. No effect of
WCC was observed on yield-scaled GHG emission in NT soil, ranging between 1.8 and
2.2 Mg CO₂ Mg⁻¹ grain rice (Table 3).

291 Discussion

The lack of effect of NT and WCC on rice grain yield has been observed in other 292 293 studies based in field experiments in Southern Brazil (Bayer et al. 2014; Zshornack et al. 2016) and in other regions of the world (Biay-Singh et al. 2008; Huang et al. 2015). The 294 295 reasonable explanation for this the no effect of NT and WCC on lack of rice yield response 296 to management is that all nutrient demand for the crop is supplied by fertilization, andIn 297 contrast to, different thanobservations occurs in Brazilian uplandsupland areas, the 298 impact of soil management practices on soil water retention and availability, soil 299 compaction and Al toxicity, among others, did was not reflected in benefits for to flooded 300 rice-cultivated under flooding. This is a very different result than that observed in uplands, where water availability for crops has been is the main variable related to theresponsible 301 302 for the increase of crop yields in rainfed no-till cropping systems (Franchini et al. 2012). 303 The strong impact of the NT system decreasing from 15 to 28% soil CH₄ emissions in comparison to the traditional CT combined with winter fallow is in 304 305 agreement to previous studies, which accounted for a decrease from of 21% in Southern 306 Brazil (Bayer et al. 2014) and from 22 to 27% in China (Zhang et al. 2015; Zhao et al. 2016). This lower CH₄ emission in NT has been attributed to the no soil incorporation of 307 biomass residues, a source of labile C readily available for methanogenic microorganisms 308 309 that are highly active in subsurface soil layers, considering that soil reduction degree is larger (<150 to 200 mV) compared to surface layers (Zschornack et al. 2011; da Silva et 310 311 al. 2011; Wang et al. 1993). 312 The above mentioned explanation for higher soil CH₄ emissions under CT than under NT is reasonable for WCC treatments, where an expressive biomass input was 313 observed (3.9-6.8 Mg ha⁻¹) but it is not to does not explain the lower CH₄ emissions in 314

NT for fallow system where biomass input by weed accounted 0.7 Mg ha⁻¹ in CT and 1.6

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Mg ha⁻¹ in NT system (Table 2). Higher CH_4 -C emissions in CT (472 kg ha⁻¹) than NT (398 kg ha⁻¹), even under winter fallow that determined a low amount of vegetable residues before soil tillage in spring, suggest that higher CH_4 -C emissions in CT soils are also related to the effect of ploughing and disking in the breakdown of soil aggregates and exposure of labile C for methanogenic microorganisms.

321 Despite the wide range of soil N₂O-N emissions (0.25-0.86 kg ha⁻¹) among soil management systems, due to high random variation, as evidenced by the high coefficient 322 323 of variation, ranging from 55 to 149%, the observed difference of seasonal N₂O emissions were not statistically significant (p > 0.05) and cannot be attributed to the management 324 325 practices (Table 1). High variation indexes are a common feature in studies on N₂O emissions in agricultural soils, that have been attributed to the temporal and spatial 326 327 variability that add up to the complex combination of the several soil variables 328 encompassed in N_2O production in soils (Hénault *et al.* 2012). The magnitude of N_2O 329 emissions is similar that observed in previous studies conducted in paddy rice fields 330 (Bayer et al. 2014; Bayer et al. 2015; Zschornack et al. 2018), but lower (for the same 331 time frame, 4 months) than usually observed in upland soils (Weller et al. 2015), which is possibly associated to the fact that emissions are restricted to the initial phase after 332 flooding and after suspension of water supply, motivated by the presence of soil nitrate 333 334 and by the activity of nitrifying and denitrifying microorganisms (Zou et al. 2005; Zschornack et al. 2011). 335

No-tillage increased SOC stocks in this lowland soil, a similar tendency observed in previous studies in Southern Brazil (Nascimento *et al.* 2009; Rosa *et al.* 2011), and other countries such as Uruguay (Terra *et al.* 2006) and China (Rui and Zhang 2010). The annual accumulation rates of SOC (0.45 to 0.65 Mg ha⁻¹) in this lowland soil under NT system were similar than those observed in upland soils considering the time of NT **Commented [A19]:** Needs to be rewritten due to English rather than content

341 adoption, clay content and soil depth (Bayer et al. 2006; Zanatta et al. 2007; Veloso et al. 342 2018). Even these results suggest a similar impact of no soil disturbance on soil organic 343 matter stabilization in lowland and upland soils, the importance of mechanisms 344 encompassed on SOC stabilization may be different. In lowland soils cultivated with irrigated rice, flooding weakens and favors the disruption of aggregates, with lower 345 346 impact of NT on physical protection of SOC (Nascimento et al. 2009). However, this effect must be partially off-set by the higher magnitude of organo-mineral interactions in 347 flooded soils (Hanke and Dick 2017). 348

Winter cover crops had no effect on SOC stocks in NT soil, despite the higher 349 350 annual biomass input in comparison to winter fallow. There is no reasonable explanation for the non-effect of WCC on SOC stocks, except for the spatial variability of the area 351 and the fact that the production of biomass in winter is quite variable and, in several years 352 353 a low crop development was observed either due to frequent rainfall and flooding of the 354 soil or poor seed quality. The increase of biomass input in upland soils results in a linear increase of SOC stocks (Bayer et al. 2006; Veloso et al. 2018), despite in other studies a 355 356 non-linear relationship has been verified, especially considering C stocks in surface layers with a low soil C saturation deficit, which determines a lower capacity of soil to 357 stabilization of added C (Corbeels et al. 2016). 358

The GHG balance in rice production systems is represented by the net GHG, which is the sum of seasonal CH₄ and N₂O emissions, and CO₂ emissions for which annual net change in SOC was used as a proxy (Piva *et al.* 2012). Across winter managements, NT-based rice production system had a net GHG lower than under CT combined with winter fallow. These results were mainly related to the atmospheric CO₂-C sequestration in soil organic matter and to the decrease of seasonal soil CH₄-C emissions. CH₄-C was the stronger component of net GHG, representing more than 85% of total **Commented [A20]:** Usually plant residues left on the soil surface are respired within a few months and lost as CO2, especially in humid or wet conditions. Only with incorporation and inputs of N and other nutrients is there a building of SOC stocks through microbial growth and the formation of humus or more resistant OM. What was the C:N ratio of the surface residues?

366 GHG emissions which is widely reported for flooded rice fields (Bayer *et al.* 2014, 2015;

367 Kim *et al.* 2012). In contrast, SOC sequestration mitigated 12% of GHG emissions, in
368 average of NT cropping systems.

Higher yield-scaled GHG emission under CT with winter fallow than NT 369 370 combined with the different winter managements indicated that NT has potential to 371 decreased 26% of GHG emissions for each 1 Mg of grain rice produced. Thus, NT constitutes a technology with potential to increase the sustainability of irrigated rice 372 373 production in subtropical ecosystems, which is usually very contested as to its environmental impact. In general, WCC had no effect on net balance of GHG and yield-374 375 scaled GHG emissions in NT soil, which is associated to the fact that although favoring soil organic carbon sequestration by increasing crop residues input, net GHG and yield-376 377 scaled GHG emissions was offset by also favoring CH₄-C emissions. These opposite 378 effects can be illustrated in the cropping system with white oat.

379

380 Conclusions

381 No-tillage mitigated net GHG emissions in comparison to conventional tillageCT in lowland soils cultivated with flooded rice, which is mainly associated to decrease of soil 382 383 CH4-C emissions and increase of soil C sequestration. In addition to maintenance of crop 384 residues on soil surface, the results suggest that decrease of CH4-C emissions in no-tillage soils are also related to the lower exposure of labile C from soil to methanogenic 385 386 microorganisms. Winter cover crops have no clear impact on net balance of GHG 387 emissions and yield-scaled GHG emissions in NT soil, mainly because soil C sequestration due to higher aboveground biomass was partially off-set by increased CH4-388 389 C emissions.

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Commented [A22]: Can you put this in the context of other Summer crops such as sorghum or maize?

Commented [A23]: And lack of soil disturbance. The maintenance of surface residues was not tested properly in this study. It would have had to have a design with paired CT-WC for each species for comparison with NT-WC.

391 Conflicts of interest

392 The authors declare no conflicts of interest.

393

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- 399

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572	Figure 1. Seasonal	emissions of (a	a) methane ((CH ₄) and (F) nitrous oxide ((N_2O) in a
572	rigure 1. Seasonai		i) memane (CI14) and (t) muous oniue	(1 v ₂ O) III a

- 573 subtropical Gleysol subjected to different soil winter management (fallow or cover
- crops) combined with conventional tillage (CT) or no-tillage (NT) in Southern Brazil.
- 575 Vertical lines denote the mean standard deviation. Different letters on the bars indicate
- 576 significant difference between treatments by Test of Skott-Knott at 5% level. $ns = \frac{no}{100}$
- 577 <u>in</u>significant.
- 578

Commented [A24]: What is the *n*? No of observations? Was it a single measurement/plot?



- 580 Figura 2. Soil organic C (SOC) contents in soil profile (a) and stocks at 0-40 cm soil
- 581 layer (b) of a Gleysol subjected to different soil winter managements (fallow or cover
- crops) combined with conventional tillage (CT) or no-tillage (NT) in southern Brazil.
- 583 Vertical lines denote the mean standard deviation $\underline{n=4?}$. Different letters on the bars
- indicate significant difference between treatments by Test of Skott-Knott at 5% level. ns
- 585 =<u>no-in</u>significant.

586 **Table 1**. Analysis of variance data (values of calculated F and P of analyzed variables)

for seasonal emissions of CH_4 e-and N_2O , stock and accumulation rate of soil organic

588 carbon (SOC) stock and accumulation rate, net balance of GHG (net GHG), rice yield

and yield-scaled GHG emissions in a subtropical Gleysol subjected to flooding rice production

590 in Southern Brazil

Dependent variable	F value	Pr > Fc
Rice Yield	1.45 ns	0.276
Seasonal CH ₄ -C	11.87 *	< 0.001
Seasonal N ₂ O-N	0.44 ns	0.776
SOC stock	5.88 *	< 0.001
SOC accumulation rate	9.25 #	< 0.001
Net GHG	2.76 #	0.077
Yield-Scaled GWP	5.34 #	0.010

591

592 * indicates significant effect by the Skott-Knott test at 5% of probability and # indicates significant effect at 10% of

593 probability; and ns indicates the lack of significant effect.

Commented [A25]: Is this according to management? Is it CT compared with NT or WCC? Need to explain this better in the legend or you could delete this table and include the data in Table 2 and the Figures

Table 2. Rice grain yield, biomass input by rice, cover crops and weeds in different

595 winter managements (fallow or cover crops) combined with conventional tillage (CT) or

596 no-tillage (NT) in a subtropical Gleysol subjected to flooding rice production in

597 Southern Brazil.

_

		Rice		Weeds / Cover crops	Total
Tillage System	Winter management	Grain Yield	Biomass	Biomass	
]	Mg ha ⁻¹	
СТ	Fallow	8.2 ns	7.7	Ó.7	10.4
NT	Fallow	8.6	10.0	1.6	11.6
	Ryegrass	8.5	9.2	5.5	14.8
	White oat	8.2	10.0	6.8	16.8
	B. trefoil	8.7	9.8	3.9	13.7

598

599 ns= no significant by analysis of variance at 5% significance level.

600 Biomass input by rice, weed and cover crops were not statistically analyzed because they are only

501 support data for interpretation of soil organic carbonSOC and greenhouse GHG gases results.

Table 3. Seasonal methane (CH₄) and nitrous oxide (N₂O) emissions, annual accumulation rate
of SOC, net balance of GHG (net GHG) and yield-scaled GHG in a flooded rice field on a Gleysol
subjected to winter managements (fallow or cover crops) combined with conventional tillage (CT)
and no-tillage (NT) in Southern Brazil.

Tillage	Winter	CU	CH ₄ N ₂ O	SOC	Net	Yield-Scaled GHG
System	management	CH ₄			GHG	
			Mg CC	D ₂ eq. ha ⁻¹ y	r ¹	Mg CO ₂ eq. Mg ⁻¹ grain
СТ	Fallow	21.4 a	0.3 a	0.0 b	21.7 a	2.6 a
NT	Fallow	18.1 b	0.1 a	1.7 a	16.5 b	1.9 b
	Ryegrass	16.3 b	0.2 a	1.6 a	14.8 c	1.8 b
	White oat	20.0 a	0.4 a	2.4 a	18.0 b	2.2 b
	B. trefoil	17.0 b	0.2 a	1.7 a	15.4 c	1.8 b

606

607 Different letters in column indicate significant significant difference (5%)